CATS RUN WATERSHED FINALTMDL Fayette County

Prepared for:

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¹Final TMDL Cats Run Watershed Fayette County, Pennsylvania

Introduction

This report presents the Total Maximum Daily Loads (TMDLs) developed for stream segments in the Cats Run Watershed (Attachment A). These were done to address the impairments noted on the 1996 Pennsylvania Section 303(d) list of impaired waters, required under the Clean Water Act, and covers one segment on this list (shown in Table 1) plus one additional segment. Depressed pH and high levels of metals caused these impairments. All impairments resulted from acid drainage from abandoned coal mines. The TMDL addresses the three primary metals associated with acid mine drainage (iron, manganese, aluminum), and pH.

	Table 1. 303(d) Sub-List							
		State Wa	ater Plan	(SWP) Su	bbasin: 19-G	Whiteley Cre	eek	
Year	Miles	Segment	DEP	Stream	Designated	Data	Source	EPA
		ID	Stream	Name	Use	Source		305(b)
			Code					Cause
								Code
1996	1.5	4917	41314	Cats Run	WWF	305(b) Report	RE	pН
1998	1.61	4917	41314	Cats Run	WWF	SWMP	AMD	pН
2000	0.44	4917	41314	Cats Run	WWF	SWMP	AMD	pН
2002	No a	additional asse	ssment	Cats Run	WWF	SWMP	AMD	pН
1996	Not	on 1996 303(d	d) List					
1998	Not	on 1998 303(d	d) List					
2000	1.16	990102-	41314	Cats Run	WWF	SWMP	AMD	Metals
		1055-TVP						& Ph
2002	No additional assessment		Cats Run	WWF	SWMP	AMD	Metals	
								& Ph

Resource Extraction=RE Warm Water Fishes=WWF Surface Water Monitoring Program = SWMP Abandoned Mine Drainage = AMD

See Attachment E, Excerpts Justifying Changes Between the 1996, 1998 and Draft 2000 Section 303(d) Lists.

The use designations for the stream segments in this TMDL can be found in PA Title 25 Chapter 93.

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¹ Pennsylvania's 1995 and 1998 Section 303(d) lists were approved by the Environmental Protection Agency (EPA). The 2000 Section 303(d) list was not required by U. S. Environmental Protection Agency. The 1996 Section 303(d) list provides the basis for measuring progress under the 1996 lawsuit settlement of *American Littoral Society and Public Interest Group of Pennsylvania v. EPA*.

Directions to the Cats Run Watershed

Cats Run is part of the Monongahela River Basin located in southwestern Pennsylvania, approximately 50 miles south of Pittsburgh. It is adjacent to the town of Masontown in Fayette County. The Cats Run basin can be accessed by traveling on U.S. Route 21 and taking State Route 166 south through Masontown. Route 166 parallels Cats Run just south of town. From Pittsburgh, take U.S. Route 79 south to U.S. Route 21 east, to State Route 166 south.

Segments addressed in this TMDL

There are no active mining operations in the watershed. All of the discharges in the watershed are from abandoned mines and will be treated as non-point sources. The distinction between non-point and point sources in this case is determined on the basis of whether or not there is a responsible party for the discharge. Where there is no responsible party the discharge is considered to be a non-point source. Each segment on the Section 303(d) list will be addressed as a separate TMDL. These TMDLs will be expressed as long-term, average loadings. Due to the nature and complexity of mining effects on the watershed, expressing the TMDL as a long-term average gives a better representation of the data used for the calculations. See Attachment D for TMDL calculations.

Clean Water Act Requirements

Section 303(d) of the 1972 Clean Water Act requires states, territories, and authorized tribes to establish water quality standards. The water quality standards identify the uses for each waterbody and the scientific criteria needed to support that use. Uses can include designations for drinking water supply, contact recreation (swimming), and aquatic life support. Minimum goals set by the Clean Water Act require that all waters be "fishable" and "swimmable."

Additionally, the federal Clean Water Act and the U.S. Environmental Protection Agency's (USEPA) implementing regulations (40 CFR Part 130) require:

- States to develop lists of impaired waters for which current pollution controls are not stringent enough to meet water quality standards (the list is used to determine which streams need TMDLs);
- States to establish priority rankings for waters on the lists based on severity of pollution and the designated use of the waterbody; states must also identify those waters for which TMDLs will be developed and a schedule for development;
- States to submit the list of waters to USEPA every two years (April 1 of the even numbered years);

- States to develop TMDLs, specifying a pollutant budget that meets state water quality standards and allocate pollutant loads among pollution sources in a watershed, e.g., point and nonpoint sources; and
- USEPA to approve or disapprove state lists and TMDLs within 30 days of final submission.

Despite these requirements, states, territories, authorized tribes, and USEPA have not developed many TMDLs since 1972. Beginning in 1986, organizations in many states filed lawsuits against the USEPA for failing to meet the TMDL requirements contained in the federal Clean Water Act and its implementing regulations. While USEPA has entered into consent agreements with the plaintiffs in several states, many lawsuits still are pending across the country.

In the cases that have been settled to date, the consent agreements require USEPA to backstop TMDL development, track TMDL development, review state monitoring programs, and fund studies on issues of concern (e.g., AMD, implementation of nonpoint source Best Management Practices (BMPs), etc.). These TMDLs were developed in partial fulfillment of the 1996 lawsuit settlement of American Littoral Society and Public Interest Group of Pennsylvania v. EPA.

Section 303(d) Listing Process

Prior to developing TMDLs for specific waterbodies, there must be sufficient data available to assess which streams are impaired and should be on the Section 303(d) list. With guidance from the USEPA, the states have developed methods for assessing the waters within their respective jurisdictions.

The primary method adopted by the Pennsylvania Department of Environmental Protection (Pa. DEP) for evaluating waters changed between the publication of the 1996 and 1998 Section 303(d) lists. Prior to 1998, data used to list streams were in a variety of formats, collected under differing protocols. Information also was gathered through the Section 305(b)² reporting process. Pa. DEP is now using the Unassessed Waters Protocol (UWP), a modification of the USEPA Rapid Bioassessment Protocol II (RPB-II), as the primary mechanism to assess Pennsylvania's waters. The UWP provides a more consistent approach to assessing Pennsylvania's streams.

The assessment method requires selecting representative stream segments based on factors such as surrounding land uses, stream characteristics, surface geology, and point source discharge locations. The biologist selects as many sites as necessary to establish an accurate assessment for a stream segment; the length of the stream segment can vary between sites. All the biological surveys included kick-screen sampling of benthic macroinvertebrates, habitat surveys, and measurements of pH, temperature, conductivity, dissolved oxygen, and alkalinity. Benthic macroinvertebrates are identified to the family level in the field.

² Section 305(b) of the Clean Water Act requires a biannual description of the water quality of the waters of the state.

After the survey is completed, the biologist determines the status of the stream segment. The decision is based on the performance of the segment using a series of biological metrics. If the stream is determined to be impaired, the source and cause of the impairment is documented. An impaired stream must be listed on the state's Section 303(d) list with the documented source and cause. A TMDL must be developed for the stream segment. A TMDL is for only one pollutant. If a stream segment is impaired by two pollutants, two TMDLs must be developed for that stream segment. In order for the process to be more effective, adjoining stream segments with the same source and cause listing are addressed collectively, and on a watershed basis.

Basic Steps for Determining a TMDL

Although all watersheds must be handled on a case-by-case basis when developing TMDLs, there are basic processes or steps that apply to all cases. They include:

- 1. Collection and summarization of pre-existing data (watershed characterization, inventory contaminant sources, determination of pollutant loads, etc.);
- 2. Calculate TMDL for the waterbody using USEPA approved methods and computer models;
- 3. Allocate pollutant loads to various sources;
- 4. Determine critical and seasonal conditions;
- 5. Submit draft report for public review and comments; and
- 6. USEPA approval of the TMDL.

Watershed History

The Cats Run basin is located in the Allegheny plateau's physiographic province and drains directly into the Monongahela River. Topography consists of gently rolling hills with maximum relief generally less than 500 feet. Elevations range from 750 feet at the confluence with the Monongahela to a maximum of just over 1200 feet.

Surface geology ranges in age from the Upper Pennsylvania to the Lower Permian found at the highest elevations. The basin contains three commercially valuable coal seams, the Pittsburgh, Sewickley, and Redstone Coals. The Pittsburgh is stratigraphically the lowest coal that crops out in the basin and is the most valuable. It underlies the majority of the basin and has been extensively deep mined throughout the early to mid twentieth century. Structurally the watershed is located between the Lambert syncline and the Uniontown anticline. The Cats Run Watershed contains a variety of land uses with a majority being forestland, crop and pastureland, and urban area (Masontown).

Abandoned deep mines into the Pittsburgh coal underlie the majority of the Cats Run Watershed. These mines operated beginning in the early 1900's through approximately 1970. The town of Masontown grew concurrently with the coal and coke industry. Coal pillars left behind to support the mine roof and the coal underclay are the main sources of pyrite needed to form acid mine drainage. Drainage from the mines has had a severe effect on surface and groundwater in

the watershed. The water quality data show acidic water with high concentrations of iron, manganese, and aluminum in all but the upper section of Cats Run.

TMDL Endpoints

One of the major components of a TMDL is the establishment of an instream numeric endpoint, which is used to evaluate the attainment of applicable water quality. An instream numeric endpoint, therefore, represents the water quality goal that is to be achieved by implementing the load reductions specified in the TMDL. The endpoint allows for comparison between observed instream conditions and conditions that are expected to restore designated uses. The endpoint is based on either the narrative or numeric criteria available in water quality standards.

Because of the nature of the pollution sources in the watershed, the TMDLs component makeup will be load allocations that are specified above a point in the stream segment. All allocations will be specified as long-term average daily concentrations. These long-term average daily concentrations are expected to meet water quality criteria 99 percent of the time. Pennsylvania Title 25 Chapter 96.3(c) specifies that the water quality standards must be met 99% of the time. The iron TMDLs are expressed at total recoverable as the iron data used for this analysis was reported as total recoverable. The following table shows the water quality criteria for the selected parameters.

Table 2. Applicable Water Quality Criteria

Parameter	Criterion Value (mg/l)	Total Recoverable/Dissolved
Aluminum (Al)	0.75	Total Recoverable
Iron (Fe)	1.50	30-day average; Total Recoverable
	0.3	Dissolved
Manganese (Mn)	1.00	Total Recoverable
pH *	6.0-9.0	N/A

^{*}The pH values shown will be used when applicable. In the case of freestone streams with little or no buffering capacity, the TMDL endpoint for pH will be the natural background water quality. These values are typically as low as 5.4 (Pennsylvania Fish and Boat Commission).

TMDL Elements (WLA, LA, MOS)

A TMDL equation consists of a wasteload allocation, load allocation and a margin of safety. The wasteload allocation is the portion of the load assigned to point sources. The load allocation is the portion of the load assigned to nonpoint sources. The margin of safety is applied to account for uncertainties in the computational process. The margin of safety may be expressed implicitly (documenting conservative processes in the computations) or explicitly (setting aside a portion of the allowable load).

Allocation Summary

These TMDLs will focus remediation efforts on the identified numerical reduction targets for each watershed. As changes occur in the watershed, the TMDLs may be re-evaluated to reflect current conditions. Table 3 presents the estimated reductions identified for all points in the watershed. Attachment D gives detailed TMDLs by segment analysis for each allocation point.

Table 3. Summary Table-Cats Run Watershed

		Mea	sured			Reduction
			ole Data	Allowa	able	Identified
Point	Parameter	Conc.	Load	LTA Conc.	Load	1401141104
, onic	' arameter	(mg/l)	(lb/day)	(mg/l)	(lb/day)	Percent
141		(IIIg/I)	(ID/day)	Cats Run Near Mout		reitein
141	Al	35.75	450.1	0.36	4.5	98
	Fe	40.00	503.6	0.40	5.0	99
	Mn	3.61	45.5	0.72	9.1	66
	Acidity	370.03	4658.3	0.00	0.0	100
	Alkalinity	0.00	0.0	0.00	0.0	100
142	rinamity	0.00	0.0	Cats Run		
112	Al	30.00	267.4	0.30	2.7	98
	Fe	9.74	86.8	0.58	5.2	90
	Mn	2.70	24.1	0.59	5.3	48
	Acidity	229.55	2046.1	0.00	0.0	100
	Alkalinity	0.00	0.0	0.00	0.0	100
1.42				M 41 CT 11 4 41	21.6	
143	A 1	42.00		Mouth of Tributary 41		00
	Al Fe	42.00	6.9 5.2	0.42 0.95	0.1	99 97
	Mn	31.50 3.40	0.6	0.95	0.1	
	Acidity	463.57	76.6	0.00	0.0	76 100
	Alkalinity	0.00	0.0	0.00	0.0	100
144	Aikaiiiity	0.00		vnstream of Confluence	rry/Tuibrotoury 41210	
144	A 1	26.83	44.4	0.27	0.4	0
	Al Fe	18.16	30.0	0.18	0.3	90
	Mn	3.10	5.1	0.56	0.9	0
	Acidity	208.39	344.5	0.00	0.0	0
	Alkalinity	0.00	0.0	0.00	0.0	0
145	rinamity	0.00	0.0	Mouth of Tributary 41	318	
143	Al	56.67	47.9	0.57	0.5	99
	Fe	20.00	16.9	1.40	1.2	93
	Mn	3.60	3.0	0.72	0.6	80
	Acidity	516.50	436.3	0.00	0.0	100
	Alkalinity	0.00	0.0	0.00	0.0	100
146			Cats Run Ur	ostream of Confluence w	/ Tributary 41318	
-	Al	21.93	93.9	0.22	0.9	97
	Fe	2.94	12.6	0.32	1.4	77
	Mn	3.20	13.7	0.64	2.7	55
	Acidity	144.49	618.9	0.00	0.0	100
	Alkalinity	0.00	0.0			
147				Cats Run Near Headwa	aters	
	Al	24.30	62.2	0.24	0.6	99
	Fe	2.79	7.1	0.22	0.6	92
	Mn	3.55	9.1	0.53	1.4	85
	Acidity	169.41	433.9	0.00	0.0	100
	Alkalinity	0.00	0.0			
148				lear Mouth of Tributary		
	Al	0.05	0.004	0.05	0.004	0
	Fe	0.65	0.1	0.65	0.1	0
	Mn	0.20	0.02	0.20	0.02	0
	Acidity	22.67	1.8	19.72	1.6	13
	Alkalinity	89.90	7.6			

Recommendations

Two primary programs that provide reasonable assurance for maintenance and improvement of water quality in the watershed are in effect. The PADEP's efforts to reclaim abandoned mine lands, coupled with its duties and responsibilities for issuing NPDES permits, will be the focal points in water quality improvement.

Additional opportunities for water quality improvement are both ongoing and anticipated. Historically, a great deal of research into mine drainage has been conducted by PADEP's Bureau of Abandoned Mine Reclamation, which administers and oversees the Abandoned Mine Reclamation Program in Pennsylvania, the United States Office of Surface Mining, the National Mine Land Reclamation Center, the National Environmental Training Laboratory, and many other agencies and individuals. Funding from EPA's 319 Grant program, and Pennsylvania's Growing Greener program have been used extensively to remedy mine drainage impacts. These many activities are expected to continue and result in water quality improvement.

The PA DEP Bureau of Mining and Reclamation administers an environmental regulatory program for all mining activities, mine subsidence regulation, mine subsidence insurance, and coal refuse disposal; conducts a program to ensure safe underground bituminous mining and protect certain structures form subsidence; administers a mining license and permit program; administers a regulatory program for the use, storage, and handling of explosives; provides for training, examination, and certification of applicants for blaster's licenses; and administers a loan program for bonding anthracite underground mines and for mine subsidence. Administers the EPA Watershed Assessment Grant Program, the Small Operator's Assistance Program (SOAP), and the Remining Operators Assistance Program (ROAP).

Reclaim PA is DEP's initiative designed to maximize reclamation of the state's quarter million acres of abandoned mineral extraction lands. Abandoned mineral extraction lands in Pennsylvania constituted a significant public liability – more than 250,000 acres of abandoned surface mines, 2,400 miles of streams polluted with mine drainage, over 7,000 orphaned and abandoned oil and gas wells, widespread subsidence problems, numerous hazardous mine openings, mine fires, abandoned structures and affected water supplies – representing as much as one third of the total problem nationally.

Mine reclamation and well plugging refers to the process of cleaning up environmental pollutants and safety hazards associated with a site and returning the land to a productive condition, similar to DEP's Brownfields program. Since the 1960's, Pennsylvania has been a national leader in establishing laws and regulations to ensure reclamation and plugging occur after active operation is completed.

Pennsylvania is striving for complete reclamation of its abandoned mines and plugging of its orphaned wells. Realizing this task is no small order, DEP has developed concepts to make abandoned mine reclamation easier. These concepts, collectively called Reclaim PA, include legislative, policy land management initiatives designed to enhance mine operator, volunteer land DEP reclamation efforts. Reclaim PA has the following four objectives.

- To encourage private and public participation in abandoned mine reclamation efforts
- To improve reclamation efficiency through better communication between reclamation partners
- To increase reclamation by reducing remining risks
- To maximize reclamation funding by expanding existing sources and exploring new sources.

Cats Run has been severely affected by acid mine drainage. The largest source being the abandoned Pittsburgh coal seam deep mine described earlier. There are three major methods of cleaning up acid mine drainage. They are remining, passive, and chemical treatment.

Remining is when an operator will remine an existing abandoned site and recover coal which could not be recovered with earlier technology. In the process, the operator removes coal and other pyretic materials which in turn reduces acid formation and accompanying metals loading to surface and ground waters. This type of activity is encouraged because it not only improves water quality and enhances land uses, but it also produces revenue for the operator. Remining techniques for this area include daylighting Pittsburgh coal deep mines and coal refuse pile reprocessing.

Passive treatment is the construction of wetlands and other treatment facilities that do not require intensive maintenance. They add alkalinity to water and/or remove metals. These types of systems can be installed if space requirements are available. They do not provide dramatic increases in water quality, but are relatively inexpensive to build and maintain.

Chemical treatment is the addition of neutralizing agents to the acid mine drainage which increases the alkalinity in the water and the pH and it also precipitates significant amount of the metals dissolved in mine drainage. This type of treatment can be very effective but it can be expensive and the treatment facilities need regular maintenance.

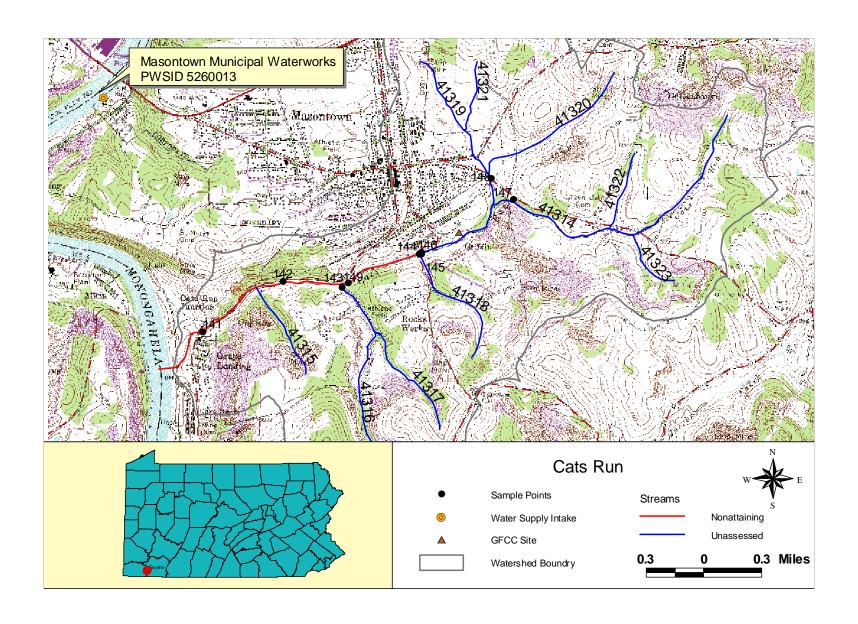
There currently is a Government Financed Construction Contract (GFCC) reclamation project underway in the Cats Run Watershed just south of Masontown that can be located on the map in Appendix A. The contract will remine any recoverable Pittsburgh coal and properly reclaim a 4-acre area consisting of abandoned highwall, spoil areas, and subsidence. The abandoned underground mine will be daylighted and backfilled. Removal of the remaining Pittsburgh crop will reduce metals loads, associated with previous mine workings, that contaminate Cats Run and its tributaries. Surface infiltration into the mine will be reduced, therefore establishing more surface flow along the operational boundary.

Public Participation

Public notice of the draft TMDL was published in the *Pennsylvania Bulletin* on December 21, 2002 and *The Herald Standard* on January 09, 2003 to foster public comment on the allowable loads calculated. A public meeting was held on January 16, 2003 at 7 pm at the American Legion Post 423, Masontown PA, to discuss the proposed TMDL.

Attachment A

Cats Run Watershed Map



Attachment B

AMD Methodology, the pH Method, and Surface Mining Control and Reclamation Act

AMD Methodology

Two approaches are used for the TMDL analysis of AMD-affected stream segments. Both of these approaches use the same statistical method for determining the instream allowable loading rate at the point of interest. The difference between the two is based on whether the pollution sources are defined as discharges that are permitted or have a responsible party, which are considered point sources. Nonpoint sources are then any pollution sources that are not point sources.

For situations where all of the impact is due to nonpoint sources, the equations shown below are applied using data for a point in the stream. The load allocation made at that point will be for all of the watershed area that is above that point. For situations where there are only point-source impacts or a combination of point and nonpoint sources, the evaluation will use the point-source data and perform a mass balance with the receiving water to determine the impact of the point source.

TMDLs and load allocations for each pollutant were determined using Monte Carlo simulation. Allocations were applied uniformly for the watershed area specified for each allocation point. For each source and pollutant, it was assumed that the observed data were log-normally distributed. Each pollutant source was evaluated separately using @Risk³ by performing 5,000 iterations to determine any required percent reduction so that the water quality criteria will be met instream at least 99 percent of the time. For each iteration, the required percent reduction is:

$$PR = \max \{0, (1-Cc/Cd)\}$$
 where (1)

PR = required percent reduction for the current iteration

Cc = criterion in mg/l

Cd = randomly generated pollutant source concentration in mg/l based on the observed data

$$Cd = RiskLognorm(Mean, Standard Deviation)$$
 where (1a)

Mean = average observed concentration Standard Deviation = standard deviation of observed data

The overall percent reduction required is the 99th percentile value of the probability distribution generated by the 5,000 iterations, so that the allowable long-term average (LTA) concentration is:

$$LTA = Mean * (1 - PR99)$$
 where (2)

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³ @Risk – Risk Analysis and Simulation Add-in for Microsoft Excel, Palisade Corporation, Newfield, NY, 1990-1997

LTA = allowable LTA source concentration in mg/l

Once the required percent reduction for each pollutant source was determined, a second series of Monte Carlo simulations were performed to determine if the cumulative loads from multiple sources allow instream water quality criteria to be met at all points at least 99 percent of the time. The second series of simulations combined the flows and loads from individual sources in a stepwise fashion, so that the level of attainment could be determined immediately downstream of each source. Where available data allowed, pollutant-source flows used were the average flows. Where data were insufficient to determine a source flow frequency distribution, the average flow derived from linear regression was used.

In general, these cumulative impact evaluations indicate that, if the percent reductions determined during the first step of the analysis are achieved, water quality criteria will be achieved at all upstream points, and no further reduction in source loadings is required.

Where a stream segment is listed on the Section 303(d) list for pH impairment, the evaluation is the same as that discussed above; the pH method is fully explained in this Attachment. An example calculation from the Swatara Creek TMDL, including detailed tabular summaries of the Monte Carlo results, is presented for the Lorberry Creek TMDL in Attachment C. Information for the TMDL analysis performed using the methodology described above is contained in the TMDLs by segment section of this report in Attachment D.

Accounting for Upstream Reductions in AMD TMDLs

In AMD TMDLs, sample points are evaluated in headwaters (most upstream) to stream mouth (most downstream) order. As the TMDL evaluation moves downstream the impact of the previous, upstream, evaluations must be considered. The following examples are from the Beaver Run AMD TMDL (2003):



In the first example BR08 is the most upstream sample point and BR02 is the next downstream sample point. The sample data, for both sample points, are evaluated using @Risk (explained above) to calculate the existing loads, allowable loads, and a percentage reduction for aluminum, iron, manganese, and acidity (when flow and parameter data are available).

Any calculated load reductions for the upstream sample point, BR08, must be accounted for in the calculated reductions at sample point BR02. To do this (see table A) the allowable load is subtracted from the existing load, for each parameter, to determine the total load reduction.

Table A	Alum.	Iron	Mang.	Acidity
BR08	(#/day)	(#/day)	(#/day)	(#/day)
existing load=	3.8	2.9	3.5	0.0
allowable load=	3.8	2.9	3.5	0.0
TOTAL LOAD REDUCTION=	0.0	0.0	0.0	0.0

In table B the Total Load Reduction BR08 is subtracted from the Existing loads at BR02 to determine the Remaining Load. The Remaining Load at BR02 has the previously calculated Allowable Loads at BR02 subtracted to determine any load reductions at sample point BR02. This results in load reductions for aluminum, iron and manganese at sample point BR02.

At sample point BR05 this same procedure is also used to account for calculated reductions at sample points BR08 and BR02. As can be seen in Tables C and D this procedure results in additional load reductions for iron, manganese and acidity at sample point BR04.

Table B. Necessary Reductions at Beaver Run BR02					
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)	
Existing Loads at BR02	13.25	38.44	21.98	6.48	
Total Load Reduction BR08	0.00	0.00	0.00	0.00	
Remaining Load (Existing Load at BR02 - BR08)	13.25	38.44	21.98	6.48	
Allowable Loads at BR02	2.91	9.23	7.03	6.48	
Percent Reduction	78.0%	76.0%	68.0%	NA	
Additional Removal Required at BR02	10.33	29.21	14.95	0.00	

At sample point BR05 (the most downstream) no additional load reductions are required, see Tables E and F.

Table C	Alum.	Iron	Mang.	Acidity
BR08 & BR02	(#/day)	(#/day)	(#/day)	(#/day)
Total Load Reduction=		29.21	14.95	0.0

Table D. Necessary Reductions at Beaver Run BR04					
	AI (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)	
Existing Loads at BR04	12.48	138.80	54.47	38.76	
Total Load Reduction BR08 & BR02	10.33	29.21	14.95	0.00	
Remaining Load (Existing Load at BBR04 - TLR Sum	2.15	109.59	39.53	38.76	
Allowable Loads at BR04	8.99	19.43	19.06	38.46	
Percent Reduction	NA	82.3%	51.8%	0.8%	
Additional Removal Required at BR04	0.00	90.16	20.46	0.29	

Table E	Alum.	Iron	Mang.	Acidity
BR08 BR02 &BR04	(#/day)	(#/day)	(#/day)	(#/day)
Total Load Reduction=	10.3	29.2	14.9	0.0

Table F. Necessary Reductions at Beaver Run BR05						
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)		
Existing Loads at BR05	0.0	31.9	22.9	4.1		
Total Load Reduction BR08, BR02 & BR04	10.3	119.4	35.4	0.3		
Remaining Load (Existing Load at BBR05 - TLR Sum	NA	NA	NA	3.8		
Allowable Loads at BR05	0.0	20.4	15.1	4.1		
Percent Reduction	NA	NA	NA	NA		
Additional Removal Required at BR05	0.0	0.0	0.0	0.0		

Although the evaluation at sample point BR05 results in no additional removal this does not mean there are no AMD problems in the stream segment BR05 to BR04. The existing and allowable loads for BR05 show that iron and manganese exceed criteria and, any abandoned mine discharges in this stream segment will be addressed.

Method for Addressing Section 303(d) Listings for pH

There has been a great deal of research conducted on the relationship between alkalinity, acidity, and pH. Research published by the Pa. Department of Environmental Protection demonstrates that by plotting net alkalinity (alkalinity-acidity) vs. pH for 794 mine sample points, the resulting pH value from a sample possessing a net alkalinity of zero is approximately equal to six (Figure 1). Where net alkalinity is positive (greater than or equal to zero), the pH range is most commonly six to eight, which is within the USEPA's acceptable range of six to nine and meets Pennsylvania water quality criteria in Chapter 93.

The pH, a measurement of hydrogen ion acidity presented as a negative logarithm, is not conducive to standard statistics. Additionally, pH does not measure latent acidity. For this reason, and based on the above information, Pennsylvania is using the following approach to address the stream impairments noted on the Section 303(d) list due to pH. The concentration of acidity in a stream is at least partially chemically dependent upon metals. For this reason, it is extremely difficult to predict the exact pH values, which would result from treatment of abandoned mine drainage. Therefore, net alkalinity will be used to evaluate pH in these TMDL calculations. This methodology assures that the standard for pH will be met because net alkalinity is a measure of the reduction of acidity. When acidity in a stream is neutralized or is restored to natural levels, pH will be acceptable. Therefore, the measured instream alkalinity at the point of evaluation in the stream will serve as the goal for reducing total acidity at that point. The methodology that is applied for alkalinity (and therefore pH) is the same as that used for other parameters such as iron, aluminum, and manganese that have numeric water quality criteria.

Each sample point used in the analysis of pH by this method must have measurements for total alkalinity and total acidity. Net alkalinity is alkalinity minus acidity, both being in units of milligrams per liter (mg/l) CaCO₃. The same statistical procedures that have been described for use in the evaluation of the metals is applied, using the average value for total alkalinity at that point as the target to specify a reduction in the acid concentration. By maintaining a net alkaline stream, the pH value will be in the range between six and eight. This method negates the need to specifically compute the pH value, which for mine waters is not a true reflection of acidity. This method assures that Pennsylvania's standard for pH is met when the acid concentration reduction is met.

There are several documented cases of streams in Pennsylvania having a natural background pH below six. If the natural pH of a stream on the Section 303(d) list can be established from its upper unaffected regions, then the pH standard will be expanded to include this natural range. The acceptable net alkalinity of the stream after treatment/abatement in its polluted segment will be the average net alkalinity established from the stream's upper, pristine reaches added to the acidity of the polluted portion in question. Summarized, if the pH in an unaffected portion of a stream is found to be naturally occurring below six, then the average net alkalinity for that portion (added to the acidity of the polluted portion) of the stream will become the criterion for the polluted portion. This "natural net alkalinity level" will be the criterion to which a 99 percent confidence level will be applied. The pH range will be varied only for streams in which a natural unaffected net alkalinity level can be established. This can only be done for streams that have upper segments that are not impacted by mining activity. All other streams will be required to reduce the acid load so the net alkalinity is greater than zero 99% of time.

Reference: Rose, Arthur W. and Charles A. Cravotta, III 1998. Geochemistry of Coal Mine Drainage. Chapter 1 in Coal Mine Drainage Prediction and Pollution Prevention in Pennsylvania. Pa. Dept. of Environmental Protection. Harrisburg. Pa.

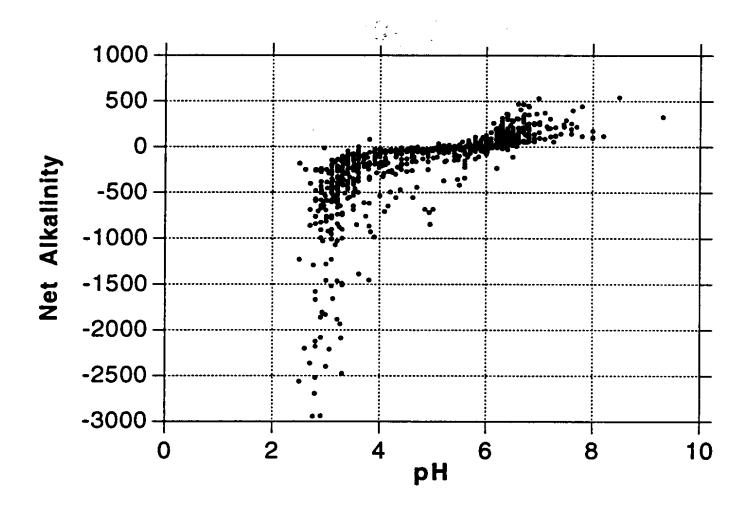


Figure 1. Net Alkalinity vs. pH. Taken from Figure 1.2 Graph C, pages 1-5, of Coal Mine Drainage Prediction and Pollution Prevention in Pennsylvania

Surface Mining Control and Reclamation Act

The Surface Mining Control and Reclamation Act of 1977 (SMCRA, Public Law 95-87) and its subsequent revisions were enacted to established a nationwide program to, among other things, protect the beneficial uses of land or water resources, and pubic health and safety from the adverse effects of current surface coal mining operations, as well as promote the reclamation of mined areas left without adequate reclamation prior to August 3, 1977. SMCRA requires a permit for the development of new, previously mined, or abandoned sites for the purpose of surface mining. Permittees are required to post a performance bond that will be sufficient to ensure the completion of reclamation requirements by the regulatory authority in the event that the applicant forfeits. Mines that ceased operating by the effective date of SMCRA, (often called "pre-law" mines) are not subject to the requirements of SMCRA.

Title IV of the Act is designed to provide assistance for reclamation and restoration of abandoned mines, while Title V states that any surface coal mining operations shall be required to meet all applicable performance standards. Some general performance standards include:

- Restoring the affected land to a condition capable of supporting the uses which it was capable of supporting prior to any mining,
- Backfilling and compacting (to insure stability or to prevent leaching of toxic materials) in order to restore the approximate original contour of the land with all highwalls being eliminated, and topsoil replaced to allow revegetation, and
- Minimizing the disturbances to the hydrologic balance and to the quality and quantity
 of water in surface and ground water systems both during and after surface coal mining
 operations and during reclamation by avoiding acid or other toxic mine drainage.

For purposes of these TMDLs, point sources are identified as NPDES-permitted discharge points, and nonpoint sources include discharges from abandoned mine lands, including but not limited to, tunnel discharges, seeps, and surface runoff. Abandoned and reclaimed mine lands were treated in the allocations as nonpoint sources because there are no NPDES permits associated with these areas. In the absence of an NPDES permit, the discharges associated with these land uses were assigned load allocations.

The decision to assign load allocations to abandoned and reclaimed mine lands does not reflect any determination by EPA as to whether there are, in fact, unpermitted point source discharges within these land uses. In addition, by establishing these TMDLs with mine drainage discharges treated as load allocations, EPA is not determining that these discharges are exempt from NPDES permitting requirements.

Related Definitions

Pre-Act (Pre-Law) - Mines that ceased operating by the effective date of SMCRA and are not subject to the requirements of SMCRA.

Bond – A instrument by which a permittee assures faithful performance of the requirements of the acts, this chapter, Chapters 87-90 and the requirements of the permit and reclamation plan.

Postmining pollution discharge – A discharge of mine drainage emanating from or hydrologically connected to the permit area, which may remain after coal mining activities have been completed, and which does not comply with the applicable effluent requirements described in Chapters 87.102, 88.92, 88.187, 88.292, 89.52 or 90.102. The term includes minimal-impact postmining discharges, as defined in Section of the Surface Mining Conservation and Reclamation Act.

Forfeited Bond – Bond money collected by the regulatory authority to complete the reclamation of a mine site when a permittee defaults on his reclamation requirements.

Attachment C

Example Calculation: Lorberry Creek

Lorberry Creek was evaluated for impairment due to high metals contents in the following manner: the analysis was completed in a stepwise manner, starting at the headwaters of the stream and moving to the mouth. The Rowe Tunnel (Swat-04) was treated as the headwaters of Lorberry Creek for the purpose of this analysis.

- 1. A simulation of the concentration data at point Swat-04 was completed. This estimated the necessary reduction needed for each metal to meet water quality criteria 99 percent of the time as a long-term average daily concentration. Appropriate concentration reductions were made for each metal.
- 2. A simulation of the concentration data at point Swat-11 was completed. It was determined that no reductions in metals concentrations are needed for Stumps Run at this time. Therefore, no TMDL for metals in Stumps Run is required at this time.
- 3. A mass balance of loading from Swat-04 and Swat-11 was completed to determine if there was any need for additional reductions as a result of combining the loads. No additional reductions were necessary.
- 4. The mass balance was expanded to include the Shadle Discharge (L-1). It was estimated that best available technology (BAT) requirements for the Shadle Discharge were adequate for iron and manganese. There is no BAT requirement for aluminum. A wasteload allocation was necessary for aluminum at point L-1.

There are no other known sources below the Shadle Discharge. However, there is additional flow from overland runoff and one unnamed tributary not impacted by mining. It is reasonable to assume that the additional flow provides assimilation capacity below point L-1, and no further analysis is needed downstream.

The calculations are detailed in the following section (Tables 1-8). Table 9 shows the allocations made on Lorberry Creek.

1. A series of four equations was used to determine if a reduction was needed at point Swat-04, and, if so the magnitude of the reduction.

	Table 1. Equations Used for Rowe Tunnel Analysis (SWAT 04)						
	Field Description	Equation	Explanation				
1	Swat-04 Initial Concentration	= Risklognorm (Mean, St Dev)	This simulates the existing concentration				
	Value (Equation 1A)		of the sampled data.				
2	Swat-04 % Reduction (from	= (Input a percentage based on	This is the percent reduction for the				
	the 99 th percentile of percent	reduction target)	discharge.				
	reduction)						
3	Swat-04 Final Concentration	= Sampled Value x (1-percent	This applies the given percent reduction				
	Value	reduction)	to the initial concentration.				
4	Swat-04 Reduction Target	= Maximum (0, 1- Cd/Cc)	This computes the necessary reduction,				
	(PR)		if needed, each time a value is sampled.				
			The final reduction target is the 99 th				
			percentile value of this computed field.				

2. The reduction target (PR) was computed taking the 99th percentile value of 5,000 iterations of the equation in row four of Table 1. The targeted percent reduction is shown, in boldface type, in the following table.

Table 2. Swat-04 Estimated Target Reductions					
Name	Swat-04 Aluminum	Swat-04 Iron	Swat-04 Manganese		
Minimum =	0	0.4836	0		
Maximum =	0.8675	0.9334	0.8762		
Mean =	0.2184	0.8101	0.4750		
Std. Deviation =	0.2204	0.0544	0.1719		
Variance =	0.0486	0.0030	0.0296		
Skewness =	0.5845	-0.8768	-0.7027		
Kurtosis =	2.0895	4.3513	3.1715		
Errors Calculated =	0	0	0		
Targeted Reduction % =	72.2	90.5	77.0		
Target #1 (Perc%)=	99	99	99		

3. This PR value was used as the percent reduction in the equation in row three of Table 1. Testing was done to see that the water quality criterion for each metal was achieved at least 99 percent of the time. This verified the estimated percent reduction necessary for each metal. Table 3 shows, in boldface type, the percent of the time criteria for each metal was achieved during 5,000 iterations of the equation in row three of Table 1.

Table 3. Swat	Table 3. Swat-04 Verification of Target Reductions					
	Swat-04	Swat-04	Swat-04			
Name	Aluminum	Iron	Manganese			
Minimum =	0.0444	0.2614	0.1394			
Maximum =	1.5282	2.0277	1.8575			
Mean =	0.2729	0.7693	0.4871			
Std Deviation =	0.1358	0.2204	0.1670			
Variance =	0.0185	0.0486	0.0279			
Skewness =	1.6229	0.8742	1.0996			
Kurtosis =	8.0010	4.3255	5.4404			
Errors Calculated =	0	0	0			
Target #1 (value) (WQ Criteria)=	0.75	1.5	1			
Target #1 (Perc%)=	99.15	99.41	99.02			

4. These same four equations were applied to point Swat-11. The result was that no reduction was needed for any of the metals. Tables 4 and 5 show the reduction targets computed for, and the verification of, reduction targets for Swat-11.

Table 4. Swat-11 Estimated Target Reductions				
	Swat-11	Swat-11	Swat-11	
Name	Aluminum	Iron	Manganese	
Minimum =	0.0000	0.0000	0.0000	
Maximum =	0.6114	0.6426	0.0000	
Mean =	0.0009	0.0009	0.0000	
Std Deviation =	0.0183	0.0186	0.0000	
Variance =	0.0003	0.0003	0.0000	
Skewness =	24.0191	23.9120	0.0000	
Kurtosis =	643.4102	641.0572	0.0000	
Errors Calculated =	0	0	0	
Targeted Reduction % =	0	0	0	
Target #1 (Perc%) =	99	99	99	

Table 5. Sw	Table 5. Swat-11 Verification of Target Reductions							
	Swat-11 Swat-11 Swat-11							
Name	Aluminum	Iron	Manganese					
Minimum =	0.0013	0.0031	0.0246					
Maximum =	1.9302	4.1971	0.3234					
Mean =	0.0842	0.1802	0.0941					
Std Deviation =	0.1104	0.2268	0.0330					
Variance =	0.0122	0.0514	0.0011					
Skewness =	5.0496	4.9424	1.0893					
Kurtosis =	48.9148	48.8124	5.1358					
Errors Calculated =	0	0	0					
WQ Criteria =	0.75	1.5	1					
% of Time Criteria Achieved =	99.63	99.60	100					

5. Table 6 shows variables used to express mass balance computations.

Table 6. Variable Descriptions for Lorberry Creek Calculations			
Description	Variable Shown		
Flow from Swat-04	Q _{swat04}		
Swat-04 Final Concentration	C_{swat04}		
Flow from Swat-11	Q _{swat11}		
Swat-11 Final Concentration	C_{swat11}		
Concentration below Stumps Run	C_{stumps}		
Flow from L-1 (Shadle Discharge)	Q_{L1}		
Final Concentration From L-1	C_{L1}		
Concentration below L-1	$C_{ m allow}$		

6. Swat-04 and Swat-11 were mass balanced in the following manner:

The majority of the sampling done at point Swat-11 was done in conjunction with point Swat-04 (20 matching sampling days). This allowed for the establishment of a significant correlation between the two flows (the R-squared value was 0.85). Swat-04 was used as the

base flow, and a regression analysis on point Swat-11 provided an equation for use as the flow from Swat-11.

The flow from Swat-04 (Q_{swat04}) was set into an @RISK function so it could be used to simulate loading into the stream. The cumulative probability function was used for this random flow selection. The flow at Swat-04 is as follows (Equation 1):

$$Q_{\text{swat04}} = \text{RiskCumul(min,max,bin range, cumulative percent of occurrence)}$$
 (1)

The RiskCumul function takes four arguments: minimum value, maximum value, the bin range from the histogram, and cumulative percent of occurrence.

The flow at Swat-11 was randomized using the equation developed through the regression analysis with point Swat-04 (Equation 2).

$$Q_{\text{swat}11} = Q_{\text{swat}}04 \times 0.142 + 0.088 \tag{2}$$

The mass balance equation is as follows (Equation 3):

$$C_{\text{stumps}} = ((Q_{\text{swat04}} * C_{\text{swat04}}) + (Q_{\text{swat11}} * C_{\text{swat11}}))/(Q_{\text{swat04}} + Q_{\text{swat11}})$$
(3)

This equation was simulated through 5,000 iterations, and the 99th percentile value of the data set was compared to the water quality criteria to determine if standards had been met. The results show there is no further reduction needed for any of the metals at either point. The simulation results are shown in Table 7.

Table 7. Verification of Meeting Water Quality Standards Below Stumps Run				
	Below Stumps	Below Stumps	Below Stumps	
Name	Run Aluminum	Run Iron	Run Manganese	
Minimum =	0.0457	0.2181	0.1362	
Maximum =	1.2918	1.7553	1.2751	
Mean =	0.2505	0.6995	0.4404	
Std Deviation =	0.1206	0.1970	0.1470	
Variance =	0.0145	0.0388	0.0216	
Skewness =	1.6043	0.8681	1.0371	
Kurtosis =	7.7226	4.2879	4.8121	
Errors Calculated =	0	0	0	
WQ Criteria =	0.75	1.5	1	
% of Time Criteria Achieved =	99.52	99.80	99.64	

7. The mass balance was expanded to determine if any reductions would be necessary at point L-1.

The Shadle Discharge originated in 1997, and very few data are available for it. The discharge will have to be treated or eliminated. It is the current site of a USGS test

remediation project. The data that were available for the discharge were collected at a point prior to a settling pond. Currently, no data for effluent from the settling pond are available.

Modeling for iron and manganese started with the BAT-required concentration value. The current effluent variability based on limited sampling was kept at its present level. There was no BAT value for aluminum, so the starting concentration for the modeling was arbitrary. The BAT values for iron and manganese are 6 mg/l and 4 mg/l, respectively. Table 8 shows the BAT-adjusted values used for point L-1.

Table 8. L-1 Adjusted BAT Concentrations					
Parameter	Measured Value BAT adjusted Value				
	Average	Standard	Average	Standard	
	Conc.	Deviation	Conc.	Deviation	
Iron	538.00	19.08	6.00	0.21	
Manganese	33.93	2.14	4.00	0.25	

The average flow (0.048 cfs) from the discharge will be used for modeling purposes. There were not any means to establish a correlation with point Swat-04.

The same set of four equations used for point Swat-04 was used for point L-1. The equation used for evaluation of point L-1 is as follows (Equation 4):

$$C_{\text{allow}} = ((Q_{\text{swat04}} * C_{\text{swat04}}) + (Q_{\text{swat11}} * C_{\text{swat11}}) + (Q_{\text{L1}} * C_{\text{L1}})) / (Q_{\text{swat04}} + Q_{\text{swat11}} + Q_{\text{L1}})$$
(4)

This equation was simulated through 5,000 iterations, and the 99th percentile value of the data set was compared to the water quality criteria to determine if standards had been met. It was estimated that an 81 percent reduction in aluminum concentration was needed for point L-1.

8. Table 9 shows the simulation results of the equation above.

Table 9. Verification of Meeting Water Quality Standards Below Point L-1				
	Below L-1	Below L-1	Below L-1	
Name	Aluminum	Iron	Manganese	
Minimum =	0.0815	0.2711	0.1520	
Maximum =	1.3189	2.2305	1.3689	
Mean =	0.3369	0.7715	0.4888	
Std Deviation =	0.1320	0.1978	0.1474	
Variance =	0.0174	0.0391	0.0217	
Skewness =	1.2259	0.8430	0.9635	
Kurtosis =	5.8475	4.6019	4.7039	
Errors Calculated =	0	0	0	
WQ Criteria=	0.75	1.5	1	
Percent of time achieved=	99.02	99.68	99.48	

9. Table 10 presents the estimated reductions needed to meet water quality standards at all points in Lorberry Creek.

	Table 10. Lorberry Creek Summary						
		Measured Sample Data		Allo	Reduction Identified		
Station	Parameter	Conc. (mg/l)	Load (lbs/day)	LTA Conc. (mg/l)	Load (lbs/day)	%	
Swat 04							
	Al	1.01	21.45	0.27	5.79	73%	
	Fe	8.55	181.45	0.77	16.33	91%	
	Mn	2.12	44.95	0.49	10.34	77%	
Swat 11							
	Al	0.08	0.24	0.08	0.24	0%	
	Fe	0.18	0.51	0.18	0.51	00%	
	Mn	0.09	0.27	0.09	0.27	00%	
L-1							
	Al	34.90	9.03	6.63	1.71	81%	
	Fe	6.00	1.55	6.00	1.55	0%	
	Mn	4.00	1.03	4.00	1.03	0%	

All values shown in this table are long-term average daily values

The TMDL for Lorberry Creek requires that a load allocation be made to the Rowe Tunnel Discharge (Swat-04) for the three metals listed, and that a wasteload allocation is made to the Shadle Discharge (L-1) for aluminum. There is no TMDL for metals required for Stumps Run (Swat-11) at this time.

Margin of Safety

For this study, the margin of safety is applied implicitly. The allowable concentrations and loadings were simulated using Monte Carlo techniques and employing the @Risk software. Other margins of safety used for this TMDL analysis include the following:

- None of the data sets were filtered by taking out extreme measurements. Because the 99 percent level of protection is designed to protect for the extreme event, it was pertinent not to filter the data set.
- Effluent variability plays a major role in determining the average value that will meet water quality criteria over the long term. This analysis maintained that the variability at each point would remain the same. The general assumption can be made that a treated discharge would be less variable than an untreated discharge. This implicitly builds in another margin of safety.

Attachment DTMDLs By Segment

CATS RUN

The TMDL for Cats Run consists of load allocations of three tributaries and five sampling sites along the stream. Following is an explanation of the TMDL for each allocation point.

Cats Run is listed for low pH from AMD as being the cause of the degradation to the stream. The method and rationale for addressing pH is contained in Attachment B.

There currently is no entry for Cats Run on the Pa Section 303(d) list for impairment due to sulfates. Although water quality data indicates high sulfate concentrations in Cats Run, no TMDL will be completed for sulfates. The nearest potable water supply intake is located approximately 2.5 miles downstream of the mouth of Cats Run at the Masontown Municipal Waterworks (PWSID 5260013), which intakes its water from the Monongahela River.

TMDL calculations- Cats Run, Sampling Point 147

The TMDL for sample point 147, located on Cats Run, consists of a load allocation to all of the area above the point shown in Attachment A. The load allocation for this point was computed using water-quality sample data collected at point 147. The average flow, measured at the sampling point 147 (0.31 MGD), is used for these computations.

There currently is no entry for this segment on the Pa 303(d) list for impairment due to pH. Sample data at point 147 shows pH ranging between 3.38 and 4.57; pH will be addressed as part of this TMDL. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

An allowable long-term average in-stream concentration was determined at point 147 for aluminum, iron, manganese, and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D1. Load Allocations at Point 147						
	Measured	Measured Sample Allowable				
	Data				Identified	
Parameter	Conc.	Load	LTA conc.	Load	%	
	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)		
	_					
Al	24.30	62.2	0.24	0.6	99	
Fe	2.79	7.1	0.22	0.6	92	
Mn	3.55	9.1	0.53	1.4	85	
Acidity	169.41	433.9	0.00	0.0	100	
Alkalinity	0.00	0.0			_	

TMDL calculations- Mouth of Tributary 41319, Sampling Point 148

The TMDL for sample point 148, located at the mouth of tributary 41319, consists of a load allocation to all of the area above the point shown in Attachment A. The load allocation for this tributary was computed using water-quality sample data collected at point 148. The average flow, measured at the sampling point 148 (0.0096 MGD), is used for these computations.

There currently is no entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 148 shows pH ranging between 6.98 and 7.77. The sample point is net alkaline and pH will not be analyzed. The method and rationale for addressing pH is contained in Attachment B.

An allowable long-term average in-stream concentration was determined at point 147 for acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D2. Load Allocations at Point 148							
	Measured	Measured Sample Allowable					
	Da	nta			Identified		
Parameter	Conc.	Load	LTA conc.	Load	%		
	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)			
Al	0.05	0.004	0.05	0.004	0		
Fe	0.65	0.1	0.65	0.1	0		
Mn	0.20	0.02	0.20	0.02	0		
Acidity	22.67	1.8	19.72	1.6	13		
Alkalinity	89.90	7.6					

<u>TMDL Calculation – Sampling Point 146 on Cats Run upstream of confluence w/ Tributary</u> 41318

The TMDL for sampling point 146, located on Cats Run, consists of a load allocation of the area between sample points 146 and 147. The load allocation for this point was computed using water-quality sample data collected at point 146. The average flow, measured at the sampling point 146 (0.51 MGD), is used for these computations.

There currently is no entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 146 shows pH ranging between 3.57 and 4.62; pH will be addressed as part of this TMDL because of the mining impacts. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

The existing and allowable loading for point 146 for all parameters was computed using water-quality sample data collected at the point. This was based on the sample data for the point and did not account for any load reductions already specified from upstream sources. The load reduction from points 148 and 147 was subtracted from the existing load at point 146 and was compared to the allowable load at 146 for each parameter to determine if any further reductions were needed at this point.

An allowable long-term average in-stream concentration was determined at point 146 for aluminum, iron, manganese, and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that

multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D3. Load Allocations at Point 146					
	Measured				
	Data		Allov	wable	
	Conc.	Load	LTAConc.	Load	
Parameter	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)	
Al	21.93	93.9	0.22	0.9	
Fe	2.94	12.6	0.32	1.4	
Mn	3.20	13.7	0.64	2.7	
Acidity	144.49 618.9		0.00	0.0	
Alkalinity	0.00	0.0			

The loading reductions for point 146, shows the total load that was removed from upstream sources. This value, for each parameter, was then subtracted from the existing load at point 146. This value was then compared to the allowable load at point 146. Reductions at point 146 are necessary for any parameter that exceeded the allowable load at this point. Table D4 shows a summary of all loads that affect point 146. Table D5 illustrates the necessary reductions at point 146. The results of this analysis show that reductions for aluminum, iron, manganese, and acidity are necessary at this point.

Table D4. Summary of All Loads that Affect Point 146						
	Al (#/day) Fe (#/day) Mn (#/day) Acidity (#/day)					
Sample Point 147						
load reduction=	61.6	6.6	7.7	433.9		
Sample Point 148						
load reduction=	0.0	0.0	0.0	0.2		

Table D5. Necessary Reductions at Sample Point 146							
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)			
Existing Loads at 146	93.9	12.6	13.7	618.9			
Total Load Reduction (147 + 148)	61.6	6.6	7.7	434.1			
Remaining Load (Existing Loads at 146-TLR Sum)	32.3	6.0	6.0	184.8			
Allowable Loads at 146	0.9	1.4	2.7	0.0			
Percent Reduction	97	77	55	100			
Additional Removal Required at 146	31.4	4.7	3.3	184.8			

The average flow, measured at sample point 146, is used for these computations. The TMDL for 146 consists of load allocations for aluminum, iron, manganese, and acidity to all of the area upstream of 146 shown in Attachment A. The percent reduction was calculated using below equation.

$$\left[1 - \left(\frac{\text{Allowable Loads at 58}}{\text{Remaining Load (Existing Loads at 58 - TLR Sum}}\right)\right] \times 100\%$$

TMDL calculations- Mouth of Tributary 41318, Sampling Point 145

The TMDL for sample point 145, located at the mouth of tributary 41318, consists of a load allocation to all of the area above the point shown in Attachment A. The load allocation for this point was computed using water-quality sample data collected at point 145. The average flow, measured at the sampling point 145 (0.10 MGD), is used for these computations.

There currently is no entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 145 shows pH ranging between 2.87 and 2.96; pH will be addressed as part of this TMDL. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

An allowable long-term average in-stream concentration was determined at point 145 for aluminum, iron, manganese, and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D6. Load Allocations at Point 145								
	Measure	d Sample	Allow	able	Reduction			
	Da	Data						
Parameter	Conc.	Load	LTA conc.	Load	%			
	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)				
Al	56.67	47.9	0.57	0.5	99			
Fe	20.00	16.9	1.40	1.2	93			
Mn	3.60	3.0	0.72	0.6	80			
Acidity	516.50	436.3	0.00	0.0	100			
Alkalinity	0.00	0.0			_			

TMDL Calculation – Sampling Point 144 on Cats Run below confluence w/ Tributary 41318

The TMDL for sampling point 144 on Cats Run consists of a load allocation of the area between points 144 and 146 shown in Attachment A. The load allocation for this stream segment was computed using water-quality sample data collected at point 144. The average flow, measured at the sampling point 144 (0.20 MGD), is used for these computations.

There currently is an entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 144 shows pH ranging between 3.13 and 4.56; pH will be addressed as part of this TMDL because of the mining impacts. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

The existing and allowable loading for point 144 for all parameters was computed using water-quality sample data collected at the point. This was based on the sample data for the point and did not account for any load reductions already specified from upstream sources. The load reduction from points 145, 146, 147, and 148 was summed and then subtracted from the existing load at point 144. This was compared to the allowable load at 144 for each parameter to determine if any further reductions were needed at this point.

An allowable long-term average in-stream concentration was determined at point 144 for aluminum, iron, manganese and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met

99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D7. Load Allocations at Point 144										
	Measured	d Sample								
	Da	ıta	Allow	Allowable						
	Conc.	Load	LTAConc.	Load						
Parameter	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)						
Al	26.83	44.4	0.27	0.4						
Fe	18.16	30.0	0.18	0.3						
Mn	3.10	5.1	0.56	0.9						
Acidity	208.39	344.5	0.00	0.0						
Alkalinity	0.00	0.0	· · · · · · · · · · · · · · · · · · ·							

The loading reductions for points 145, 146, 147, and 148 were summed to show the total load that was removed from upstream sources. This value, for each parameter, was then subtracted from the existing load at point 144. This value was then compared to the allowable load at point 144. Reductions at point 144 are necessary for any parameter that exceeded the allowable load at this point. Table D8 shows a summary of all loads that affect point 144. Table D9 illustrates the necessary reductions at point 144. The results of this analysis show that reduction for iron is necessary at this point.

Table D8. Sur	Table D8. Summary of All Loads that Affect Point 144									
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)						
Sample Point 148										
load reduction=	0.0	0.0	0.0	0.2						
Sample Point 147										
load reduction=	61.6	6.6	7.7	433.9						
Sample Point 146										
load reduction=	31.4	4.7	3.3	184.8						
Sample Point 145										
load reduction=	47.4	15.7	2.4	436.3						

Table D9. Necessary Reduction	ns at Samp	ole Point 1	44	
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)
Existing Loads at 144	44.4	30.0	5.1	344.5
Total Load Reduction (Sum of 145, 146, 147, and 148)	140.4	27.0	13.4	1055.2
Remaining Load (Existing Loads at 144-TLR Sum)	NA	3.0	NA	NA
Allowable Loads at 144	0.4	0.3	0.9	0.0
Percent Reduction	NA	90	NA	NA
Additional Removal Required at 144	NA	2.7	NA	NA

The average flow, measured at sample point 144, is used for these computations. The TMDL for 144 consists of load allocations for iron to all of the area upstream of 144 shown in Attachment A. The percent reduction was calculated using below equation.

No additional loading reductions were necessary for aluminum, manganese, and acidity.

TMDL Calculation – Mouth of Tributary 41316, Sampling Point 143

The TMDL for sample point 143, located at the mouth of tributary 41316, consists of a load allocation to all of the area above the point shown in Attachment A. The load allocation for this tributary was computed using water-quality sample data collected at point 143. The average flow, measured at the sampling point 143(0.020 MGD), is used for these computations.

There currently is not an entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 143 shows pH ranging between 2.84 and 3.92; pH will be addressed as part of this TMDL. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

An allowable long-term average in-stream concentration was determined at point 143 for aluminum, iron, manganese, and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and

standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D10. Load Allocations at Point 143										
	Measured Da	d Sample ata	Allov	Reduction Identified						
Parameter	Conc. Load (lbs/day)		LTA Conc. (mg/l)	Load (lbs/day)	%					
Al	42.00	6.9	0.42	0.1	99					
Fe	31.50	5.2	0.95	0.2	97					
Mn	3.40	0.6	0.82	0.1	76					
Acidity	463.57 76.6		0.00	0.0	100					
Alkalinity	0.00	0.0								

TMDL Calculation – Sampling Point 142 on Cats Run below Tributary 41316

The TMDL for sampling point 142 on Cats Run consists of a load allocation of the area between points 142 and 144 shown in Attachment A. The load allocation for this stream segment was computed using water-quality sample data collected at point 142. The average flow, measured at the sampling point 142 (1.07 MGD), is used for these computations.

There currently is an entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 142 shows pH ranging between 3.16 and 3.53; pH will be addressed as part of this TMDL because of the mining impacts. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

The existing and allowable loading for point 142 for all parameters was computed using water-quality sample data collected at the point. This was based on the sample data for the point and did not account for any load reductions already specified from upstream sources. The load reduction from points 143, 144, 145, 146, 147, and 148 was summed and then subtracted from the existing load at point 142. This was compared to the allowable load at 142 for each parameter to determine if any further reductions were needed at this point.

An allowable long-term average in-stream concentration was determined at point 142 for aluminum, iron, manganese and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D	Table D11. Load Allocations at Point 142										
	Measured	d Sample									
	Da	ıta	Allow								
	Conc.	Load	LTAConc.	Load							
Parameter	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)							
Al	30.00	267.4	0.30	2.7							
Fe	9.74	86.8	0.58	5.2							
Mn	2.70	24.1	0.59	5.3							
Acidity	229.55	2046.1	0.00	0.0							
Alkalinity	0.00	0.0									

The loading reductions for points 143, 144, 145, 146, 147, and 148 were summed to show the total load that was removed from upstream sources. This value, for each parameter, was then subtracted from the existing load at point 142. This value was then compared to the allowable load at point 142. Reductions at point 142 are necessary for any parameter that exceeded the allowable load at this point. Table D12 shows a summary of all loads that affect point 142. Table D13 illustrates the necessary reductions at point 142. The results of this analysis show that reduction for aluminum, iron, manganese, and acidity is necessary at this point.

Table D12. Su	mmary of A	all Loads th	at Affect Po	int 142
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)
Sample Point 148				
load reduction=	0.0	0.0	0.0	0.2
Sample Point 147				
load reduction=	61.6	6.6	7.7	433.9
Sample Point 146				
load reduction=	31.4	4.7	3.3	184.8
Sample Point 145				
load reduction=	47.4	15.7	2.4	436.3
Sample Point 144				
load reduction=	NA	2.7	NA	NA
Sample Point 143				
load reduction=	6.8	5.0	0.5	76.6

Table D13. Necessary Reductio	Table D13. Necessary Reductions at Sample Point 142										
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)							
Existing Loads at 142	267.4	86.8	24.1	2046.1							
Total Load Reduction (Sum of 143, 144, 145, 146, 147, and 148)	147.2	34.7	13.9	1131.8							
Remaining Load (Existing Loads at 142-TLR Sum)	120.2	52.1	10.2	914.3							
Allowable Loads at 142	2.7	5.2	5.3	0.0							
Percent Reduction	98	90	48	100							
Additional Removal Required at 142	117.5	46.9	4.9	914.3							

The average flow, measured at sample point 142, is used for these computations. The TMDL for 142 consists of load allocations for aluminum, iron, manganese, and acidity to all of the area upstream of 142 shown in Attachment A. The percent reduction was calculated using below equation.

$$\left[1 - \left(\frac{\text{Allowable Loads at 57}}{\text{Remaining Load (Existing Loads at 57 - TLR Sum}}\right)\right] \times 100\%$$

TMDL Calculation – Sampling Point 141 on Cats Run near the mouth

The TMDL for sampling point 141 on Cats Run consists of a load allocation of the area between points 141 and 142 shown in Attachment A. The load allocation for this stream segment was computed using water-quality sample data collected at point 141. The average flow, measured at the sampling point 141(1.51 MGD), is used for these computations.

There currently is an entry for this segment on the Pa Section 303(d) list for impairment due to pH. Sample data at point 141 shows pH ranging between 3.02 and 3.16; pH will be addressed as part of this TMDL because of the mining impacts. The objective is to reduce acid loading to the stream, which will in turn raise the pH and keep a net alkalinity above zero, 99% of the time. The result of this analysis is an acid loading reduction that equates to meeting standards for pH (see TMDL Endpoint section in the report, Table 2). The method and rationale for addressing pH is contained in Attachment B.

The existing and allowable loading for point 141 for all parameters was computed using water-quality sample data collected at the point. This was based on the sample data for the point and did not account for any load reductions already specified from upstream sources. The load reduction from points 142, 143, 144, 145, 146, 147, and 148 was summed and then subtracted from the existing load at point 141. This was compared to the allowable load at 141 for each parameter to determine if any further reductions were needed at this point.

An allowable long-term average in-stream concentration was determined at point 141 for aluminum, iron, manganese and acidity. The analysis is designed to produce an average value that, when met, will be protective of the water-quality criterion for that parameter 99% of the time. An analysis was performed using Monte Carlo simulation to determine the necessary long-term average concentration needed to attain water-quality criteria 99% of the time. The simulation was run assuming the data set was lognormally distributed. Using the mean and standard deviation of the data set, 5000 iterations of sampling were completed, and compared against the water-quality criterion for that parameter. For each sampling event a percent reduction was calculated, if necessary, to meet water-quality criteria. A second simulation that multiplied the percent reduction times the sampled value was run to insure that criteria were met 99% of the time. The mean value from this data set represents the long-term average concentration that needs to be met to achieve water-quality standards. The following table shows the load allocations for this stream segment.

Table D14. Load Allocations at Point 141										
	Measured	d Sample								
	Da	ıta	Allow							
	Conc.	Load	LTAConc.	Load						
Parameter	(mg/l)	(lbs/day)	(mg/l)	(lbs/day)						
Al	35.75	450.1	0.36	4.5						
Fe	40.00	503.6	0.40	5.0						
Mn	3.61	45.5	0.72	9.1						
Acidity	370.03	4658.3	0.00	0.0						
Alkalinity	0.00	0.0								

The loading reductions for points 142, 143, 144, 145, 146, 147, and 148 were summed to show the total load that was removed from upstream sources. This value, for each parameter, was then subtracted from the existing load at point 141. This value was then compared to the allowable load at point 141. Reductions at point 141 are necessary for any parameter that exceeded the allowable load at this point. Table D15 shows a summary of all loads that affect point 141. Table D16 illustrates the necessary reductions at point 141. The results of this analysis show that reduction for aluminum, iron, manganese, and acidity is necessary at this point.

Table D15. Su	mmary of A	All Loads th	at Affect Po	int 141
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)
Sample Point 148				
load reduction=	0.0	0.0	0.0	0.2
Sample Point 147				
load reduction=	61.6	6.6	7.7	433.9
Sample Point 146				
load reduction=	31.4	4.7	3.3	184.8
Sample Point 145				
load reduction=	47.4	15.7	2.4	436.3
Sample Point 144				
load reduction=	NA	2.7	NA	NA
Sample Point 143				
load reduction=	6.8	5.0	0.5	76.6
Sample Point 142				
load reduction=	117.5	46.9	4.9	914.3

Table D16. Necessary Reductio	ns at Sam	ple Point	141	
	Al (#/day)	Fe (#/day)	Mn (#/day)	Acidity (#/day)
Existing Loads at 141	450.1	503.6	45.5	4658.3
Total Load Reduction (Sum of 142, 143, 144, 145, 146, 147, and 148)	264.7	81.6	18.8	2046.1
Remaining Load (Existing Loads at 142-TLR Sum)	185.4	422.0	26.7	2612.2
Allowable Loads at 142	4.5	5.0	9.1	0.0
Percent Reduction	98	99	66	100
Additional Removal Required at 142	180.9	417	17.61	2612.2

The average flow, measured at sample point 142, is used for these computations. The TMDL for 142 consists of load allocations for aluminum, iron, manganese, and acidity to all of the area upstream of 142 shown in Attachment A. The percent reduction was calculated using below equation.

$$\left[1 - \left(\frac{\text{Allowable Loads at 57}}{\text{Remaining Load (Existing Loads at 57 - TLR Sum}}\right)\right] \times 100\%$$

Margin of Safety

PADEP used an implicit MOS in these TMDLs derived from the Monte Carlo statistical analysis. The Water Quality standard states that water quality criteria must be met at least 99% of the time. All of the @Risk analyses results surpass the minimum 99% level of protection. Another margin of safety used for this TMDL analysis results from:

- Effluent variability plays a major role in determining the average value that will meet water-quality criteria over the long-term. The value that provides this variability in our analysis is the standard deviation of the dataset. The simulation results are based on this variability and the existing stream conditions (an uncontrolled system). The general assumption can be made that a controlled system (one that is controlling and stabilizing the pollution load) would be less variable than an uncontrolled system. This implicitly builds in a margin of safety.
- A MOS is also the fact that the calculations were performed with a daily Iron average instead of the 30-day average.

Seasonal Variation

Seasonal variation is implicitly accounted for in these TMDLs because the data used represents all seasons.

Critical Conditions

The reductions specified in this TMDL apply at all flow conditions. A critical flow condition could not be identified from the data used for this analysis.

Attachment E

Excerpts Justifying Changes Between the 1996, 1998, and Draft 2002 Section 303(d) Lists

The following are excerpts from the Pennsylvania DEP Section 303(d) narratives that justify changes in listings between the 1996, 1998, and draft 2002 lists. The Section 303(d) listing process has undergone an evolution in Pennsylvania since the development of the 1996 list.

In the 1996 Section 303(d) list narrative, strategies were outlined for changes to the listing process. Suggestions included, but were not limited to, a migration to a Global Information System (GIS), improved monitoring and assessment, and greater public input.

The migration to a GIS was implemented prior to the development of the 1998 Section 303(d) list. As a result of additional sampling and the migration to the GIS some of the information appearing on the 1996 list differed from the 1998 list. Most common changes included:

- 1. mileage differences due to recalculation of segment length by the GIS;
- 2. slight changes in source(s)/cause(s) due to new EPA codes;
- 3. changes to source(s)/cause(s), and/or miles due to revised assessments;
- 4. corrections of misnamed streams or streams placed in inappropriate SWP subbasins; and
- 5. unnamed tributaries no longer identified as such and placed under the named watershed listing.

Prior to 1998, segment lengths were computed using a map wheel and calculator. The segment lengths listed on the 1998 Section 303(d) list were calculated automatically by the GIS (ArcInfo) using a constant projection and map units (meters) for each watershed. Segment lengths originally calculated by using a map wheel and those calculated by the GIS did not always match closely. This was the case even when physical identifiers (e.g., tributary confluence and road crossings) matching the original segment descriptions were used to define segments on digital quad maps. This occurred to some extent with all segments, but was most noticeable in segments with the greatest potential for human errors using a map wheel for calculating the original segment lengths (e.g., long stream segments or entire basins).

The most notable difference between the 1998 and Draft 2000 Section 303(d) lists are the listing of unnamed tributaries in 2000. In 1998, the GIS stream layer was coded to the named stream level so there was no way to identify the unnamed tributary records. As a result, the unnamed tributaries were listed as part of the first downstream named stream. The GIS stream coverage used to generate the 2000 list had the unnamed tributaries coded with the DEP's five-digit stream code. As a result, the unnamed tributary records are now split out as separate records on the 2000 Section 303(d) list. This is the reason for the change in the appearance of the list and the noticeable increase in the number of pages. After due consideration of comments from EPA and PADEP on the Draft 2000 Section 303(d) list, the Draft 2002 Pa Section 303(d) list was written in a manner similar to the 1998 Section 303(d) list.

Attachment F Water Quality Data Used In TMDL Calculations

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
5C	141	000626-1330-xrf	1023	3.11	481		0.5	42	57	4.3	1911		
81 E	141	000930-1515-cam,jm	419	3.02	388		4.5	35.00	47.00	3.65	1420		
85E	141	010125-1200-ddk,mw	570	3.25	270		14.00	27.00	31.00	3.00	508		
30G	141	010418-1200-ddk,dr	2181	3.16	341		3	39	25	3.5	1031		
Mean	141		1048	3.14	370		5.5	35.75	40.00	3.61	1217	N39°50.033'	W79°55.073'
Stdev	141		798	0.10	88		5.9	6.50	14.65	0.54	594		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
53G	142	010125-1235-ddk,mw	308	3.53	163		10.00	21.00	6.92	2.10	672		
99G	142	010418-1230-ddk,dr	1642	3.23	286		2	38	14	3.3	865		
76 E	142	000930-1530-ddk,rxs	277	3.16	240		3	31.00	8.30	2.70	933		
Mean	142		742.19	3.31	229.55		5.00	30.00	9.74	2.70	823	N39°50.267'	W79°54.614'
Stdev	142		779.04	0.20	62.33		4.36	8.54	3.75	0.60	135.33		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
45C	143	000626-1400-xrf	20	2.86	564		2	44	35	3.4	1350		
83 E	143	000930-1600-jm,cam	27	2.84	463		0	43.00	25.00	3.80	1699		
58G	143	010125-1310-ddk,mw	5	2.93	463		2.25	47	33	3.3	876		
43G	143	010418-1300-ddk,dr	3	3.92	365		33	34	33	3.1	1249		
Mean	143		14	3.14	464		9.3	42.00	31.50	3.40	1293	N39°50.257'	W79°54.260'
Stdev	143		12	0.52	81		15.8	5.60	4.43	0.29	339		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
43C	144	000626-1500-xrf	172	3.13	309		0.5	38	6.6	3.8	1262		
64 E	144	000930-1700-ddk,rxs	94	3.14	258		0	33.00	47.00	3.30	1087		
59G	144	010125-?-ddk,mw	147	4.56	58	0.31	0.00	9.50	0.89	2.20	443		
Mean	144		138	3.61	208	0	0.2	26.83	18.16	3.10	931	N39°50.405'	W79°53.822'
Stdev	144		40	0.82	133		0.3	15.22	25.14	0.82	431		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
47C	145	000626-1535-xrf	42	2.87	596		4.75	54	20	3.7	1975		
52G	145	010125-ddk,mw	5	2.96	503		1.50	61.00	19.00	3.00	2175		
42G	145	010418-1430-ddk,dr	164	2.95	451		0.5	55	21	4.1	1307		
Mean	145		70	2.93	516		2.2	56.67	20.00	3.60	1819	N39°50.406'	W79°53.805'
Stdev	145		83	0.05	73		2.2	3.79	1.00	0.56	455		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
52C	146	000626-1600-xrf	89	3.69	159		0	23	1.6	3.5	620		
72 E	146	000930-1645-cam,jm	82	3.70	147		0	23.00	2.50	3.60	704		
22H	146	010418-1400-ddk,dr	899	3.57	216		5.5	33	6.9	3.4	592		
51G	146		frozen?	4.62	57	0.23	4.50	8.70	0.76	2.30	429		
Mean	146		357	3.90	144	0	2.5	21.93	2.94	3.20	586	N39°50.412'	W79°53.805'
Stdev	146		470	0.49	66		2.9	10.00	2.73	0.61	115		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
59C	147	000626-1700-xrf	44	3.39	202		0	27	0.94	4.7	636		
80 E	147	000930-1730-ddk,rxs	23	3.69	149		0.5	22.00	0.36	3.60	621		
31G	147	010125-1525-ddk,mw	112	4.57	52	0.16	0.00	8.20	0.25	2.20	378		
17G	147	010418-1445-ddk,dr	674	3.38	275		2	40	9.6	3.7	617		
Mean	147		213	3.76	169	0	0.6	24.30	2.79	3.55	563	N39°50.658'	W79°53.281'
Stdev	147		310	0.56	94		0.9	13.14	4.55	1.03	124		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
58C	148	000626-1640-xrf	7	7.77	11	86	9	0.02	0.58	0.12	48		
69 E	148	000930-1730-jm,cam	2	7.39	47	107	2	0.04	0.54	0.31	57		
56G	148	010125-1455-ddk,mw	11	6.98	10	76	8.50	0.08	0.82	0.16	65		
Mean	148		7	7.38	23	90	6.5	0.05	0.65	0.20	57	N39°50.753'	W79°53.410'
Stdev	148		5	0.40	21	16	3.9	0.03	0.15	0.10	9		

Bottle ID	Site	date-time-samplerID	Flow (gpm)	рН	Acidity (mg/L)	Alk (mg/L)	TSS (mg/L)	Al (ppm)	Fe (ppm)	Mn (ppm)	SO4 (ppm)	Latitude	Longitude
82 E	149	000930-1600-ddk,rxs	6	2.67	846		1	99.00	60.00	4.85	2378		
49G	149	010125-1340-ddk,mw	3	2.78	778		0.50	84.00	57.00	4.50	2252		
Mean	149		5	2.73	812		0.7	91.50	58.50	4.68	2315	N39°50.260'	W79°54.231'
Stdev	149		2	0.08	48		0.4	10.61	2.12	0.25	89		

Attachment GComment and Response

The following comments were submitted by the United States Environmental Protection Agency Region 3, on February 03, 2003 in regards to the proposed TMDL for Cats Run.

1. Note that any remining permit with National Pollutant Discharge Elimination System (NPDES) permitted discharges will need to conform to this TMDL.

There are no remining NPDES permitted discharges in the Cats Run watershed.

2. Please consider adding text to Attachment D describing the location of the sampling point. Generally, referring to the map is adequate but sometimes the location may be ambiguous, e.g., point 143. Point 143 is presumed to be on tributary 41316 but, based on the map, could be on the mainstem. Please consider that future users of the report may only have black and white copies and may not be able to identify the locations easily.

Text was added to the report to clarify the locations of the sampling points.